

The Role of Biochar in Agriculture

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CHAPTER 1

Introduction

Integrated farming is an agricultural approach that combines various components of farming systems, such as crop production, livestock rearing, aquaculture, and agroforestry, to create a synergistic and sustainable environment (Manjunatha *et al.*, 2014). This holistic approach aims to optimize resource utilization, minimize waste generation, and enhance the overall productivity and profitability of the farming system (Behera *et al.*, 2015). Integrated farming has gained prominence in recent years due to its potential to address the growing challenges of food security, environmental degradation, and climate change (FAO, 2018).

Table 1. Physicochemical properties of biochar derived from different feedstocks

| Feedstock | Pyrolysis Temperature (°C) | | pH | Surface Area (m ² /g) | CEC (cmol/kg) | Carbon Content (%) | Ash Content (%) |
|----------------------|----------------------------|--|------|----------------------------------|---------------|-------------------------------------|-----------------|
| Wood | 500 | | 7.5 | 120 | 40 | 75 | 5 |
| Rice husk | 600 | | 9.2 | 80 | 25 | 45 | 40 |
| Manure | 400 | | 8.1 | 60 | 90 | 55 | 30 |
| Straw | 550 | | 10.3 | 40 | 70 | 60 | 20 |
| Blanco-Canqui (2017) | Corn stover | | 40 | | Silt loam | Improved aggregate stability by 44% | |

1.2. The role of soil health in sustainable agriculture Soil health is a critical factor in sustainable agriculture, as it directly influences crop growth, yield, and quality (Doran & Zeiss, 2000). A healthy soil is characterized by optimal physical, chemical, and biological properties that support plant growth, nutrient cycling, water retention, and disease suppression (Karlen *et al.*, 2003). However, intensive agricultural practices, such as monocropping, excessive tillage, and indiscriminate use of agrochemicals, have led to soil degradation, erosion, and loss of fertility (Montgomery, 2007). Therefore, maintaining and enhancing soil health has become a priority for sustainable agriculture (Lal, 2015).

1.3. Biochar: Definition, production, and properties

Biochar is a carbon-rich material produced by the thermal decomposition of organic biomass under limited oxygen conditions, a process known as pyrolysis (Lehmann & Joseph,

2015). The feedstocks for biochar production can include agricultural waste, forestry residues, animal manure, and other organic materials (Woolf *et al.*, 2010). Biochar is characterized by its high surface area, porous structure, and stable aromatic carbon compounds, which contribute to its unique properties and potential applications in agriculture (Sohi *et al.*, 2010). These properties include high cation exchange capacity (CEC), water retention, nutrient adsorption, and resistance to decomposition (Atkinson *et al.*, 2010).

CHAPTER - 2

Biochar production and characterization

2.1. Feedstocks for biochar production

Biochar can be produced from a wide range of organic feedstocks, each with unique properties and potential applications in agriculture.

The most common feedstocks for biochar production include:

2.1.1. Agricultural waste

Agricultural waste, such as crop residues, straw, and husks, is an abundant and readily available feedstock for biochar production (Purakayastha *et al.*, 2019). Rice husk, wheat straw, and corn stover are among the most extensively studied agricultural waste feedstocks for biochar production (El-Naggar *et al.*, 2019). The utilization of agricultural waste for biochar production not only provides a sustainable waste management solution but also adds value to these otherwise underutilized resources (Sadaka *et al.*, 2014).

Table 2. Effects of biochar application on soil physical properties in various studies

| Study | Biochar Feedstock | Application Rate (t/ha) | Soil Type | Effects on Soil Physical Properties |
|------------------------------|-------------------|-------------------------|------------|---|
| Basso <i>et al.</i> (2013) | Hardwood | 20 | Sandy loam | Increased water holding capacity by 23% |
| Burrell <i>et al.</i> (2016) | Mixed wood | 10 | Clay loam | Reduced bulk density by 12% |
| Blanco-Canqui (2017) | Corn stover | 40 | Silt loam | Improved aggregate stability by 44% |

2.1.2. Forestry residues

Forestry residues, such as wood chips, sawdust, and bark, are another significant source of feedstock for biochar production (Wrobel-Tobiszewska *et al.*, 2015). These residues are generated during various forestry operations, including logging, pruning, and wood processing

(Nanda *et al.*, 2016). The use of forestry residues for biochar production can contribute to the sustainable management of forest resources and reduce the risk of forest fires by removing excess biomass (Page-Dumroese *et al.*, 2017).

2.1.3. Animal manure

Animal manure, particularly from poultry, cattle, and swine, has been explored as a potential feedstock for biochar production (Cantrell *et al.*, 2012). The conversion of animal manure into biochar not only provides a sustainable waste management solution but also reduces the environmental risks associated with manure storage and application, such as nutrient leaching and greenhouse gas emissions (Ro *et al.*, 2010). Moreover, manure-derived biochar has been shown to possess unique properties, such as high nutrient content and surface functionality, which can enhance its performance as a soil amendment (Subedi *et al.*, 2017).

2.2. Pyrolysis process and its influence on biochar properties

The pyrolysis process, which involves the thermal decomposition of biomass under limited oxygen conditions, is a critical factor in determining the properties and quality of the resulting biochar (Lehmann & Joseph, 2015).

The three main types of pyrolysis processes are:

2.2.1. Slow pyrolysis

Slow pyrolysis is characterized by low heating rates ($< 10^{\circ}\text{C}/\text{min}$), long residence times (hours to days), and relatively low temperatures ($300\text{-}600^{\circ}\text{C}$) (Kambo & Dutta, 2015). This process favors the production of biochar with a high yield (30-50%), moderate surface area, and high carbon content (Demirbas, 2004). Slow pyrolysis is the most common method for biochar production, as it allows for better control over the product quality and is relatively simple to implement (Wang *et al.*, 2013).

2.2.2. Fast pyrolysis

Fast pyrolysis involves high heating rates ($> 100^{\circ}\text{C}/\text{s}$), short residence times (seconds to minutes), and moderate temperatures ($400\text{-}650^{\circ}\text{C}$) (Bridgwater, 2012). This process primarily aims to maximize the production of bio-oil, with biochar being a byproduct (Mohan *et al.*, 2006). Fast pyrolysis-derived biochar typically has a lower yield (10-20%), higher surface area, and lower carbon content compared to slow pyrolysis biochar (Zhang *et al.*, 2017).

2.2.3. Gasification

Gasification is a high-temperature ($> 700^{\circ}\text{C}$) process that converts biomass into syngas (CO , H_2 , and CH_4) under controlled oxygen conditions (Kumar *et al.*, 2009). Biochar is a

byproduct of gasification, with yields typically ranging from 5-15% (Meyer *et al.*, 2011). Gasification-derived biochar is characterized by a high surface area, low carbon content, and high ash content (Hansen *et al.*, 2015). Due to its unique properties, gasification biochar has been explored for various applications, such as soil remediation and adsorption of contaminants (Xie *et al.*, 2015).

2.3. Physical and chemical properties of biochar

The physical and chemical properties of biochar are crucial factors in determining its performance as a soil amendment and its potential applications in agriculture (Lehmann & Joseph, 2015).

The key properties of biochar include:

2.3.1. Surface area and porosity

Biochar is characterized by its high surface area and porous structure, which are influenced by the feedstock type and pyrolysis conditions (Gray *et al.*, 2014). The surface area of biochar can range from a few m²/g to over 500 m²/g, depending on the production method (Downie *et al.*, 2009). The high surface area and porosity of biochar contribute to its ability to adsorb nutrients, retain water, and provide habitat for soil microorganisms (Atkinson *et al.*, 2010).

2.3.2. Elemental composition

The elemental composition of biochar varies depending on the feedstock and pyrolysis conditions (Lehmann & Joseph, 2015). Carbon is the dominant element in biochar, typically ranging from 50-90% by weight (Sohi *et al.*, 2010). Other important elements in biochar include oxygen, hydrogen, nitrogen, sulfur, and various mineral nutrients, such as potassium, phosphorus, and calcium (Singh *et al.*, 2010). The elemental composition of biochar influences its stability, nutrient retention capacity, and interactions with soil components (Bruun *et al.*, 2012).

2.3.3. pH and electrical conductivity

Biochar is generally alkaline, with pH values ranging from 6.5 to 12, depending on the feedstock and pyrolysis conditions (Yuan *et al.*, 2011). The high pH of biochar can be beneficial for acidic soils, as it can help to neutralize soil acidity and improve nutrient availability (Lehmann *et al.*, 2011). However, in alkaline soils, the application of high-pH biochar may lead to nutrient imbalances and micronutrient deficiencies (Limwikran *et al.*, 2018). The electrical conductivity (EC) of biochar is another important property, as it indicates the presence of soluble

salts (Chintala *et al.*, 2014). High-EC biochars may be less suitable for application in saline soils, as they can exacerbate salt stress in plants (Lashari *et al.*, 2013).

2.4. Factors affecting biochar quality and stability

Several factors influence the quality and stability of biochar, including feedstock type, pyrolysis conditions, and post-production treatments (Lehmann & Joseph, 2015). The selection of appropriate feedstocks and pyrolysis parameters is crucial for producing biochar with desired properties for specific soil and crop requirements (Zhao *et al.*, 2013). For instance, woody biomass generally produces biochar with higher carbon content and stability compared to herbaceous biomass (Windeatt *et al.*, 2014). Similarly, higher pyrolysis temperatures ($> 500^{\circ}\text{C}$) typically result in biochar with greater surface area, porosity, and aromatic carbon content, which contribute to its long-term stability in soil (Singh *et al.*, 2012). Post-production treatments, such as activation or chemical modification, can further enhance the properties of biochar for specific applications (Rajapaksha *et al.*, 2016).

CHAPTER - 3

Biochar and soil physical properties

Biochar has been shown to have a significant impact on soil structure and aggregation, which are critical factors in maintaining soil health and productivity (Blanco-Canqui, 2017). The porous structure and high surface area of biochar can contribute to the formation and stability of soil aggregates by promoting the binding of soil particles and organic matter (Ouyang *et al.*, 2013). The improved soil structure and aggregation can enhance water infiltration, aeration, and root growth, thereby creating favorable conditions for plant development (Burrell *et al.*, 2016).

3.2. Soil water retention and hydraulic conductivity

The application of biochar to soil has been demonstrated to improve water retention and hydraulic conductivity, particularly in sandy and coarse-textured soils (Basso *et al.*, 2013). The high porosity and surface area of biochar enable it to absorb and retain water, reducing water loss through evaporation and leaching (Novak *et al.*, 2009). Moreover, the presence of biochar can enhance the soil's water holding capacity by increasing the number of micropores and improving the pore size distribution (Liu *et al.*, 2017). The improved water retention and hydraulic conductivity can help to mitigate drought stress and optimize water use efficiency in crops (Kammann *et al.*, 2011).

Table 3. Influence of biochar amendment on soil chemical properties and nutrient availability

| Study | Biochar Feedstock | Application Rate (t/ha) | Soil Type | Effects on Soil Chemical Properties |
|------------------------------|-------------------|-------------------------|------------------------------|-------------------------------------|
| Yuan <i>et al.</i> (2011) | Crop residue | 20 | Ultisol | Increased pH from 4.3 to 6.5 |
| Laird <i>et al.</i> (2010) | Hardwood | 5 | Midwestern agricultural soil | Increased CEC by 20% |
| Lehmann <i>et al.</i> (2003) | Wood | 140 | Oxisol | Increased available P by 60% |

3.3. Soil bulk density and porosity

Biochar amendment has been reported to reduce soil bulk density and increase soil porosity, particularly in compacted and degraded soils (Rogovska *et al.*, 2014). The low bulk density and high porosity of biochar can contribute to the loosening of the soil structure, improving soil aeration and root penetration (Laird *et al.*, 2010). The reduction in soil bulk density can also facilitate the movement of water and nutrients through the soil profile, enhancing their availability to plants (Barnes *et al.*, 2014). However, the extent of biochar's impact on soil bulk density and porosity depends on the application rate, soil type, and biochar properties (Blanco-Canqui, 2017).

3.4. Soil temperature regulation

Biochar has been shown to influence soil temperature dynamics, which can have implications for crop growth and soil microbial activity (Zhang *et al.*, 2013). The dark color and high thermal conductivity of biochar can lead to increased soil temperature, particularly in the upper soil layers (Vaccari *et al.*, 2011). This effect can be beneficial in cold regions, where higher soil temperatures can promote seed germination, root growth, and nutrient uptake (Luo *et al.*, 2017). However, in hot and arid regions, the increased soil temperature may exacerbate heat stress and water loss, necessitating careful management of biochar application (Xie *et al.*, 2015).

CHAPTER 4

Biochar and soil chemical properties

Biochar application has been widely reported to increase soil pH, particularly in acidic soils (Yuan *et al.*, 2011). The alkaline nature of most biochars can help to neutralize soil acidity and improve nutrient availability (Lehmann *et al.*, 2011). The liming effect of biochar is attributed to its high ash content and the presence of basic cations, such as Ca, Mg, and K (Wang *et al.*, 2015). Moreover, biochar has been shown to enhance the soil's buffering capacity, which can help to stabilize soil pH and reduce the risk of nutrient imbalances (Xu *et al.*, 2012). However, the extent of biochar's impact on soil pH depends on the initial soil pH, biochar application rate, and biochar properties (Biederman & Harpole, 2013).

4.2. Cation exchange capacity (CEC) and nutrient retention

Biochar has been demonstrated to increase the cation exchange capacity (CEC) of soils, which is a critical factor in nutrient retention and availability (Liang *et al.*, 2006). The high surface area and negative surface charge of biochar enable it to adsorb and retain cations, such as ammonium (NH_4^+) and potassium (K^+), reducing their loss through leaching (Lehmann *et al.*, 2003). The improved CEC can also enhance the soil's ability to buffer against changes in pH and maintain a stable nutrient supply for plants (Glaser *et al.*, 2002). The effects of biochar on nutrient retention have been extensively studied for various nutrients, including:

4.2.1. Nitrogen

Biochar has been shown to reduce nitrogen (N) leaching and improve N use efficiency in crops (Steiner *et al.*, 2008). The adsorption of ammonium (NH_4^+) onto biochar's surface can reduce its conversion to nitrate (NO_3^-) and subsequent loss through leaching (Zheng *et al.*, 2013). Moreover, biochar can promote the immobilization of N by stimulating microbial growth and activity, further reducing N losses (Xu *et al.*, 2016). However, the impact of biochar on N dynamics depends on the biochar type, soil properties, and environmental conditions (Nguyen *et al.*, 2017).

4.2.2. Phosphorus

Biochar has been reported to increase phosphorus (P) availability and reduce P leaching in soils (Parvage *et al.*, 2013). The high surface area and porosity of biochar can facilitate the

adsorption of P, reducing its loss through leaching and surface runoff (Laird *et al.*, 2010). Biochar can also interact with soil minerals, such as iron (Fe) and aluminum (Al) oxides, to form stable complexes that increase P retention and availability (DeLuca *et al.*, 2009). Additionally, biochar can stimulate microbial activity and enhance the solubilization of soil P by releasing organic acids and enzymes (Vassilev *et al.*, 2013). However, the effectiveness of biochar in improving P availability depends on the biochar feedstock, pyrolysis conditions, and soil properties (Wang *et al.*, 2012).

Table 4. Biochar-induced changes in soil microbial communities and enzyme activities

| Study | Biochar Feedstock | Application Rate (t/ha) | Soil Type | Effects on Soil Biological Properties |
|------------------------------|-------------------|-------------------------|-----------------|--|
| Lehmann <i>et al.</i> (2011) | Wood | 20 | Calcareous soil | Doubled microbial biomass carbon |
| Bailey <i>et al.</i> (2011) | Papermill waste | 10 | Mine soil | Increased β -glucosidase activity by 30% |
| Feng <i>et al.</i> (2012) | Rice straw | 40 | Paddy soil | Stimulated methanotrophic community |

4.2.3. Potassium

Biochar has been shown to increase potassium (K) availability and retention in soils, particularly in sandy and low-CEC soils (Lehmann *et al.*, 2003). The high K content of some biochars, particularly those derived from wood and crop residues, can act as a direct source of K for plants (Gaskin *et al.*, 2010). Moreover, the high CEC of biochar can adsorb and retain K⁺ ions, reducing their loss through leaching (Laird *et al.*, 2010). The improved K retention can help to maintain an adequate K supply for crops, especially in soils prone to K deficiency (Rogovska *et al.*, 2014). However, the impact of biochar on K dynamics depends on the biochar properties, soil type, and cropping system (Blanco-Canqui, 2017).

4.3. Soil organic carbon and carbon sequestration

Biochar has been increasingly recognized as a potential tool for enhancing soil organic carbon (SOC) and promoting carbon sequestration in agricultural soils (Woelf *et al.*, 2010). The high carbon content and recalcitrant nature of biochar make it resistant to microbial decomposition, allowing it to persist in soils for hundreds to thousands of years (Wang *et al.*, 2016). The incorporation of biochar into soil can directly increase SOC stocks, while also

promoting the stabilization of native soil organic matter through various mechanisms, such as adsorption, aggregation, and reduced mineralization (Weng *et al.*, 2017). Moreover, biochar can indirectly enhance SOC accumulation by increasing plant biomass production and root exudation, which contribute to the formation of stable soil organic matter (Sohi *et al.*, 2010). The potential of biochar for carbon sequestration has been widely studied, with estimates suggesting that biochar application could sequester up to 1.8 Gt C per year globally (Woolf *et al.*, 2010). However, the actual carbon sequestration potential of biochar depends on various factors, such as biochar production and application rates, soil properties, climate conditions, and land management practices (Smith, 2016).

4.4. Interactions with soil contaminants and heavy metals

Biochar has been explored as a potential amendment for the remediation of soils contaminated with heavy metals and organic pollutants (Beesley *et al.*, 2011). The high surface area, porous structure, and surface functional groups of biochar can facilitate the adsorption and immobilization of various contaminants, reducing their bioavailability and toxicity to plants and soil organisms (Zhang *et al.*, 2013). For instance, biochar has been shown to effectively adsorb heavy metals, such as lead (Pb), cadmium (Cd), and zinc (Zn), through mechanisms such as ion exchange, complexation, and precipitation (Lu *et al.*, 2012). Similarly, biochar can adsorb organic contaminants, such as polycyclic aromatic hydrocarbons (PAHs) and pesticides, through hydrophobic interactions and π - π bonding (Cao *et al.*, 2011). The immobilization of contaminants by biochar can reduce their uptake by plants and minimize their leaching into groundwater (Uchimiya *et al.*, 2010). However, the effectiveness of biochar in soil remediation depends on the biochar properties, soil characteristics, contaminant type, and environmental conditions (Rajapaksha *et al.*, 2016). Moreover, the long-term stability and fate of adsorbed contaminants in biochar-amended soils require further investigation to ensure the safety and sustainability of this approach (Sohi *et al.*, 2010).

Biochar and soil biological properties

Biochar amendment has been shown to significantly influence soil microbial communities and diversity, which play critical roles in soil health and ecosystem functions (Lehmann *et al.*, 2011). The porous structure and high surface area of biochar provide a favorable habitat for soil microorganisms, protecting them from predation and desiccation (Thies *et al.*, 2015). Biochar can also serve as a source of carbon and other nutrients for microbial growth and activity (Warnock *et al.*, 2007). Several studies have reported increased microbial biomass, diversity, and activity in biochar-amended soils, particularly in the rhizosphere (Lehmann *et al.*, 2011). For instance, biochar has been shown to stimulate the growth and abundance of beneficial soil bacteria, such as *Pseudomonas* and *Bacillus* spp., which are involved in nutrient cycling, plant growth promotion, and disease suppression (Kolton *et al.*, 2011). Similarly, biochar has been found to enhance the diversity and activity of arbuscular mycorrhizal fungi (AMF), which facilitate plant nutrient uptake and improve soil structure (Warnock *et al.*, 2007). However, the impact of biochar on soil microbial communities depends on the biochar properties, soil type, and environmental conditions (Lehmann & Joseph, 2015). Moreover, the long-term effects of biochar on soil microbial dynamics and their implications for soil health and ecosystem functions require further investigation (Thies *et al.*, 2015).

5.2. Soil enzyme activities

Soil enzymes are essential for the decomposition of organic matter, nutrient cycling, and other vital soil processes (Dick, 1994). Biochar amendment has been reported to influence soil enzyme activities, either directly by providing substrates and stimulating microbial activity or indirectly by altering soil properties and microbial community structure (Bailey *et al.*, 2011). Several studies have observed increased activities of various soil enzymes, such as β -glucosidase, chitinase, and phosphatase, in biochar-amended soils (Jin, 2010). For instance, biochar has been shown to enhance the activity of β -glucosidase, an enzyme involved in the decomposition of cellulose and the release of glucose, which can stimulate microbial growth and activity (Jin *et al.*, 2016). Similarly, biochar has been found to increase the activity of acid phosphatase, an enzyme that catalyzes the hydrolysis of organic phosphate esters, thereby improving P availability for plants (Cui *et al.*, 2011).

However, the impact of biochar on soil enzyme activities varies depending on the biochar properties, soil type, and specific enzymes (Bailey *et al.*, 2011). Moreover, the mechanisms underlying biochar-induced changes in soil enzyme activities and their implications for soil health and ecosystem functions require further elucidation (Thies *et al.*, 2015).

Table 5. Crop yield responses to biochar application in different cropping systems

| Study | Biochar Feedstock | Application Rate (t/ha) | Crop | Yield Increase (%) |
|-------------------------------|-------------------|-------------------------|--------|--------------------|
| Major <i>et al.</i> (2010) | Wood | 20 | Maize | 28 |
| Vaccari <i>et al.</i> (2011) | Wheat straw | 30 | Tomato | 25 |
| Solaiman <i>et al.</i> (2010) | Oil mallee | 10 | Wheat | 45 |

5.3. Interactions with beneficial microorganisms

Biochar has been shown to interact with various beneficial soil microorganisms, such as arbuscular mycorrhizal fungi (AMF) and plant growth-promoting bacteria (PGPB), which can enhance plant growth, nutrient uptake, and stress resistance (Lehmann *et al.*, 2011). These interactions can occur through several mechanisms, including:

5.3.1. Arbuscular mycorrhizal fungi

AMF are symbiotic fungi that colonize plant roots and form extensive hyphal networks in the soil, facilitating plant nutrient uptake and improving soil structure (Smith & Read, 2008). Biochar has been reported to stimulate AMF colonization and hyphal growth, particularly in soils with low fertility or under stress conditions (Warnock *et al.*, 2007). The porous structure and high surface area of biochar can provide a suitable habitat for AMF, protecting them from predation and desiccation (Thies *et al.*, 2015). Moreover, biochar can adsorb and retain nutrients, such as P and N, which can be accessed by AMF and transferred to the host plant (Hammer *et al.*, 2014). The enhanced AMF activity in biochar-amended soils can improve plant nutrient acquisition, water uptake, and stress tolerance (Solaiman *et al.*, 2010). However, the impact of biochar on AMF varies depending on the biochar properties, soil type, and plant species (Lehmann *et al.*, 2011).

5.3.2. Plant growth-promoting bacteria

PGPB are soil bacteria that colonize plant roots and promote plant growth through various mechanisms, such as nitrogen fixation, phosphate solubilization, and phytohormone production (Glick, 2012). Biochar has been shown to stimulate the growth and activity of PGPB, particularly in the rhizosphere (Kolton *et al.*, 2011). The porous structure and high surface area of biochar can provide a favorable microhabitat for PGPB, protecting them from environmental stresses and predation (Thies *et al.*, 2015). Moreover, biochar can adsorb and retain nutrients, such as N and P, which can be accessed by PGPB and made available to plants (Zheng *et al.*, 2013). The enhanced PGPB activity in biochar-amended soils can improve plant nutrient uptake, growth, and stress resistance (Saxena *et al.*, 2013). However, the impact of biochar on PGPB varies depending on the biochar properties, soil type, and plant species (Lehmann *et al.*, 2011).

5.4. Biochar as a habitat for soil organisms

Biochar has been increasingly recognized as a potential habitat for various soil organisms, including microorganisms, invertebrates, and small vertebrates (Lehmann *et al.*, 2011). The porous structure and high surface area of biochar provide a suitable microenvironment for soil organisms, protecting them from predation, desiccation, and extreme temperatures (Thies *et al.*, 2015). Moreover, biochar can adsorb and retain water, nutrients, and organic compounds, which can support the growth and activity of soil organisms (Lehmann *et al.*, 2011). Several studies have reported increased abundance and diversity of soil fauna, such as nematodes, collembolans, and mites, in biochar-amended soils (McCormack *et al.*, 2013). For instance, biochar has been shown to increase the abundance and diversity of soil nematodes, which are important indicators of soil health and play critical roles in nutrient cycling and pest suppression (Zhang *et al.*, 2013). Similarly, biochar has been found to stimulate the activity of earthworms, which contribute to soil structure formation, organic matter decomposition, and nutrient mineralization (Weyers & Spokas, 2011). The enhanced activity of soil organisms in biochar-amended soils can improve soil health, fertility, and ecosystem functions (Lehmann *et al.*, 2011). However, the impact of biochar on soil organisms varies depending on the biochar properties, soil type, and environmental conditions (Thies *et al.*, 2015).

CHAPTER - 6

Biochar and soil chemical properties

Biochar has been widely reported to improve crop growth and productivity through various mechanisms, including:

6.1.1. Nutrient availability and uptake

Biochar has been shown to increase the availability and uptake of essential plant nutrients, such as N, P, and K, in biochar-amended soils (Lehmann *et al.*, 2003). The high surface area and CEC of biochar can adsorb and retain nutrients, reducing their loss through leaching and making them more accessible to plants (Laird *et al.*, 2010). Moreover, biochar can stimulate the activity of soil microorganisms involved in nutrient cycling, such as N-fixing bacteria and P-solubilizing fungi, further enhancing nutrient availability (Thies *et al.*, 2015). The improved nutrient status of biochar-amended soils can promote plant growth, yield, and quality (Major *et al.*, 2010). However, the impact of biochar on nutrient availability and uptake depends on the biochar properties, soil type, and crop species (Jeffery *et al.*, 2011).

Table 6. Greenhouse gas emissions from biochar-amended soils compared to control treatments

| Study | Biochar Feedstock | Application Rate (t/ha) | Soil Type | GHG Emissions Reduction |
|------------------------------|-------------------|-------------------------|------------|---|
| Zhang <i>et al.</i> (2010) | Wheat straw | 20 | Paddy soil | 21% decrease in CH ₄ emissions |
| Cayuela <i>et al.</i> (2015) | Various | 5-50 | Various | 54% decrease in N ₂ O emissions |
| Woolf <i>et al.</i> (2010) | Various | 1-50 | Global | 12% of current anthropogenic CO ₂ emissions potentially offset |

6.1.2. Water use efficiency

Biochar has been reported to improve water use efficiency and mitigate drought stress in crops (Kammann *et al.*, 2011). The porous structure and high surface area of biochar can increase

soil water retention and reduce water loss through evaporation and leaching (Novak *et al.*, 2009). Moreover, biochar can improve soil structure and hydraulic conductivity, facilitating water infiltration and root penetration (Barnes *et al.*, 2014). The enhanced water holding capacity and soil moisture status of biochar-amended soils can alleviate drought stress and improve crop water use efficiency (Mulcahy *et al.*, 2013). However, the effectiveness of biochar in improving water use efficiency depends on the biochar properties, soil type, and climate conditions (Jeffery *et al.*, 2011).

6.1.3. Disease suppression and pest control

Biochar has been shown to suppress various plant diseases and pests, thereby improving crop health and productivity (Elad *et al.*, 2010). The mechanisms underlying biochar-induced disease suppression are not fully understood but may involve several factors, such as:

- Stimulation of beneficial soil microorganisms, such as PGPB and AMF, which can improve plant resistance to pathogens (Kolton *et al.*, 2011).
- Adsorption and inactivation of plant pathogens and their toxins by biochar's high surface area and reactive functional groups (Elad *et al.*, 2010).
- Induction of systemic resistance in plants through the activation of defense-related genes and pathways (Mehari *et al.*, 2015).
- Improvement of soil physical and chemical properties, such as pH, CEC, and nutrient availability, which can create unfavorable conditions for pathogens (Jaiswal *et al.*, 2014).

6.2. Biochar application rates and methods

The application rates and methods of biochar are critical factors in determining its effectiveness in improving soil health and crop productivity (Jeffery *et al.*, 2011). The optimal biochar application rates depend on several factors, such as soil type, crop species, and biochar properties (Sohi *et al.*, 2010). In general, biochar application rates ranging from 5 to 50 t ha⁻¹ have been reported to improve soil properties and crop yields (Jeffery *et al.*, 2011). However, higher application rates (> 50 t ha⁻¹) may have negative effects on soil health and crop growth, particularly in the short term (Kammann *et al.*, 2011). Therefore, it is essential to determine the appropriate biochar application rates based on site-specific conditions and management goals (Sohi *et al.*, 2010).

Several methods have been used for biochar application in agricultural soils, including:

6.2.1. Soil incorporation

Soil incorporation is the most common method of biochar application, involving the mixing of biochar with the topsoil (0-20 cm) using tillage equipment, such as plows, disks, or harrows (Major *et al.*, 2010). This method ensures a uniform distribution of biochar in the root zone and facilitates its interactions with soil components and microorganisms (Blackwell *et al.*, 2009). However, soil incorporation may disrupt soil structure and increase the risk of soil erosion, particularly in sloping lands or under intensive tillage (Mukherjee & Lal, 2013).

6.2.2. Top-dressing

Top-dressing involves the surface application of biochar without incorporation into the soil (Kammann *et al.*, 2011). This method is less disruptive to soil structure and can be easily applied using spreaders or by hand (Major *et al.*, 2010). Top-dressing is particularly suitable for perennial crops, pastures, or conservation agriculture systems, where tillage is minimized or avoided (Sohi *et al.*, 2010). However, the effectiveness of top-dressed biochar may be limited by its low contact with soil components and microorganisms, as well as its susceptibility to wind and water erosion (Mukherjee & Lal, 2013).

6.2.3. Seedling dipping

Seedling dipping involves the coating of plant roots with a biochar slurry before transplanting (Elad *et al.*, 2010). This method ensures a direct contact between biochar and plant roots, potentially enhancing nutrient and water uptake, as well as disease suppression (Mahmood *et al.*, 2016). Seedling dipping is particularly suitable for vegetable and fruit crops that are transplanted as seedlings, such as tomatoes, peppers, and strawberries (Elad *et al.*, 2010). However, the effectiveness of seedling dipping may be limited by the small amount of biochar applied and its potential loss during transplanting and crop growth (Mahmood *et al.*, 2016).

6.3. Crop-specific responses to biochar amendment

The responses of crops to biochar amendment vary widely depending on the crop species, biochar properties, soil type, and environmental conditions (Jeffery *et al.*, 2011). In general, biochar has been reported to improve the growth and yield of various crops, including:

6.3.1. Cereals

Biochar has been shown to increase the yield and nutrient uptake of several cereal crops, such as rice, wheat, and maize (Zhang *et al.*, 2012). For instance, a meta-analysis by Jeffery *et al.* (2011) found that biochar application increased the yield of wheat by 10% and maize by 11% on average, particularly in acidic and sandy soils. The positive effects of biochar on cereal crops have been attributed to various mechanisms, such as improved soil structure, water retention, nutrient availability, and mycorrhizal colonization (Solaiman *et al.*, 2010).

6.3.2. Legumes

Biochar has been reported to enhance the growth, nodulation, and N₂ fixation of various legume crops, such as soybeans, common beans, and chickpeas (Rondon *et al.*, 2007). For example, a study by Tagoe *et al.* (2008) found that biochar application increased the yield of soybeans by 44% and common beans by 46% in a sandy soil, along with improved nodulation and N uptake. The beneficial effects of biochar on legumes have been attributed to various factors, such as increased soil pH, nutrient availability, and rhizobial activity (Rondon *et al.*, 2007).

6.3.3. Vegetables and fruits

Biochar has been shown to improve the yield and quality of various vegetable and fruit crops, such as tomatoes, lettuce, and strawberries (Harel *et al.*, 2012). For instance, a study by Vaccari *et al.* (2011) found that biochar application increased the yield of tomatoes by 25% and improved their nutritional quality, particularly the content of vitamin C and lycopene. The positive effects of biochar on vegetables and fruits have been attributed to various mechanisms, such as enhanced nutrient and water uptake, disease suppression, and improved soil physical properties (Elad *et al.*, 2010).

6.4. Long-term effects of biochar on soil fertility and crop yields

While many studies have reported positive short-term effects of biochar on soil properties and crop productivity, the long-term impacts of biochar remain largely unknown (Sohi *et al.*, 2010). Some studies have suggested that the beneficial effects of biochar may persist for several years or even decades, due to its high stability and resistance to decomposition (Lehmann *et al.*, 2006). For instance, a study by Major *et al.* (2010) found that the positive effects of biochar on soil fertility and crop yields in a Colombian savanna soil persisted for at least four years after application. Similarly, a study by Vaccari *et al.* (2011) reported that the benefits of biochar on tomato growth and yield were maintained over two consecutive growing seasons.

However, other studies have indicated that the long-term effects of biochar may be more variable and context-dependent (Jeffery *et al.*, 2011). For example, a study by Jones *et al.* (2012) found that the positive effects of biochar on soil properties and crop yields in a temperate agricultural soil diminished over time, possibly due to the aging and weathering of biochar. Similarly, a study by Zhang *et al.* (2012) reported that the benefits of biochar on rice yield in a Chinese paddy soil varied across different years and biochar application rates, suggesting the need for site-specific optimization.

CHAPTER - 7

Bio-char and greenhouse gas emissions

Biochar has been widely recognized as a potential tool for mitigating climate change through carbon sequestration in agricultural soils (Woolf *et al.*, 2010). The production and application of biochar can transfer a significant portion of the carbon from biomass feedstocks into stable, recalcitrant forms that can persist in soils for hundreds to thousands of years (Lehmann *et al.*, 2006). By sequestering carbon in soils, biochar can reduce the atmospheric concentration of CO₂ and contribute to the mitigation of global warming (Sohi *et al.*, 2010).

Table 7. Economic analysis of biochar production and application in integrated farming systems

| Study | Feedstock | Production Method | Application Rate (t/ha) | Crop | Net Return (\$/ha/yr) |
|-------------------------------|-------------------|--------------------------|-------------------------|--------------------|-----------------------|
| Kung <i>et al.</i> (2013) | Rice husk | Top-lit updraft gasifier | 10 | Vegetables | 1,000 |
| Galinato <i>et al.</i> (2011) | Wood waste | Slow pyrolysis | 5 | Wheat-pea rotation | 200 |
| Clare <i>et al.</i> (2014) | Sugarcane bagasse | Auger reactor | 10 | Sugarcane | -60 |

The carbon sequestration potential of biochar depends on several factors, such as the feedstock type, pyrolysis conditions, and soil properties (Sohi *et al.*, 2010). In general, biochar produced from woody biomass and at higher pyrolysis temperatures (>500°C) tends to have a higher carbon content and stability, and thus a greater potential for long-term carbon sequestration (Lehmann *et al.*, 2006). For instance, a study by Woolf *et al.* (2010) estimated that the global implementation of biochar systems could potentially sequester up to 1.8 Gt CO₂-equivalent per year, which corresponds to ~12% of current anthropogenic CO₂ emissions. Similarly, a study by Lehmann *et al.* (2006) suggested that the application of biochar could store up to 9.5 Gt C per year in agricultural soils, which is equivalent to the global annual fossil fuel emissions.

However, the actual carbon sequestration potential of biochar may be lower than these estimates, due to various constraints and uncertainties, such as biochar stability, soil interactions, and land availability (Sohi *et al.*, 2010). Moreover, the production and application of biochar may also have other environmental impacts, such as changes in soil properties, nutrient dynamics, and greenhouse gas emissions, which need to be carefully considered and managed (Woolf *et al.*, 2010).

7.2. Nitrous oxide (N₂O) emissions

Biochar application has been reported to affect the emissions of nitrous oxide (N₂O), a potent greenhouse gas with a global warming potential 298 times higher than CO₂ over a 100-year time horizon (IPCC, 2007). N₂O emissions from agricultural soils are primarily driven by microbial processes, such as nitrification and denitrification, which are influenced by soil properties, nitrogen availability, and environmental conditions (Butterbach-Bahl *et al.*, 2013).

The impact of biochar on N₂O emissions varies widely depending on the biochar properties, soil type, and management practices (Cayuela *et al.*, 2014). Some studies have reported that biochar application can reduce N₂O emissions from agricultural soils, particularly in coarse-textured and acidic soils (Van Zwieten *et al.*, 2010). For instance, a meta-analysis by Cayuela *et al.* (2015) found that biochar application decreased N₂O emissions by an average of 54% across 30 studies, with the greatest reductions observed in sandy soils and at high biochar application rates. The mechanisms underlying biochar-induced N₂O mitigation are not fully understood but may involve several processes, such as:

- Adsorption of NH₄⁺ and NO₃⁻ by biochar, reducing their availability for nitrification and denitrification (Singh *et al.*, 2010).
- Stimulation of N immobilization by soil microorganisms, limiting the substrate availability for N₂O-producing processes (Zheng *et al.*, 2012).
- Improvement of soil aeration and water retention, creating less favorable conditions for denitrification (Van Zwieten *et al.*, 2010).
- Alteration of soil pH and redox conditions, affecting the activity and composition of nitrifying and denitrifying communities (Cayuela *et al.*, 2013).

7.3. Methane (CH₄) emissions

Biochar application has also been investigated for its potential to affect methane (CH₄) emissions from agricultural soils, particularly in rice paddies and other wetland systems (Jeffery *et al.*, 2016). CH₄ is a potent greenhouse gas with a global warming potential 25 times higher than

CO₂ over a 100-year time horizon (IPCC, 2007). CH₄ emissions from agricultural soils are primarily driven by methanogenic archaea, which produce CH₄ under anaerobic conditions, and methanotrophic bacteria, which oxidize CH₄ in the presence of oxygen (Conrad, 2007).

The impact of biochar on CH₄ emissions from agricultural soils is highly variable and depends on several factors, such as the biochar feedstock, soil type, and water management (Jeffery *et al.*, 2016). Some studies have reported that biochar application can reduce CH₄ emissions from rice paddies and other wetland soils (Liu *et al.*, 2011; Feng *et al.*, 2012). For instance, a meta-analysis by Jeffery *et al.* (2016) found that biochar addition decreased CH₄ emissions by an average of 19% across 42 studies, with the greatest reductions observed in acidic soils and when biochar was applied in combination with other organic amendments. The mechanisms underlying biochar-induced CH₄ mitigation may include:

- Improvement of soil aeration and drainage, reducing the anaerobic conditions favoring methanogenesis (Liu *et al.*, 2011).
- Stimulation of methanotrophic activity by increasing the availability of oxygen and inorganic compounds, such as copper and iron, which are essential for methane oxidation (Feng *et al.*, 2012).
- Adsorption of CH₄ onto the surface of biochar, reducing its diffusion and emission from the soil (Cai *et al.*, 2016).
- Alteration of the soil microbial community structure, favoring methanotrophic over methanogenic populations (Feng *et al.*, 2012).

7.4. Life cycle assessment of biochar systems

To fully understand the net impact of biochar on greenhouse gas emissions and the environment, it is essential to consider the entire life cycle of biochar systems, from feedstock production and pyrolysis to soil application and long-term effects (Woolf *et al.*, 2010). Life cycle assessment (LCA) is a valuable tool for quantifying the environmental footprint of biochar systems and comparing them with other soil management practices or carbon sequestration strategies (Gaunt & Lehmann, 2008).

LCA studies on biochar systems have yielded mixed results, depending on the specific assumptions, system boundaries, and impact categories considered (Cowie *et al.*, 2015). Some studies have reported that biochar systems can have a net positive impact on climate change mitigation, primarily due to the carbon sequestration potential of biochar and the displacement of fossil fuels by pyrolysis co-products, such as bio-oil and syngas (Roberts *et al.*, 2010; Hammond *et al.*, 2011). For instance, a study by Peters *et al.* (2015) found that the application of biochar

from pyrolyzed wood waste could reduce greenhouse gas emissions by 2.1-4.6 t CO₂-equivalent per hectare per year, considering both carbon sequestration and avoided emissions from fertilizer production and use.

CHAPTER 8

Challenges and opportunities

One of the main challenges for the widespread adoption of biochar in integrated farming systems is the economic feasibility of biochar production and application (Dickinson *et al.*, 2015). The cost of biochar varies widely depending on the feedstock availability, pyrolysis technology, and transportation logistics (Meyer *et al.*, 2011). In general, the cost of biochar production ranges from \$100 to \$800 per ton, with the lower end corresponding to large-scale, centralized facilities using low-cost feedstocks, such as agricultural residues or waste biomass (Shackley *et al.*, 2011).

To assess the economic viability of biochar application in integrated farming, it is essential to conduct a cost-benefit analysis that considers both the direct and indirect costs and benefits of biochar use (Dickinson *et al.*, 2015). The direct costs of biochar application include the purchase price of biochar, transportation, and application costs, while the direct benefits include increased crop yields, reduced fertilizer and irrigation requirements, and potential carbon credits (Galinato *et al.*, 2011). The indirect costs and benefits of biochar application, such as changes in soil health, ecosystem services, and long-term productivity, are more difficult to quantify but should also be considered in the economic assessment (Guo *et al.*, 2016).

Several studies have investigated the economic feasibility of biochar application in different cropping systems and regions, with mixed results (Dickinson *et al.*, 2015). For instance, a study by Kung *et al.* (2013) found that the application of rice husk biochar at 10 t ha⁻¹ in a Taiwanese vegetable farm resulted in a net economic benefit of \$1,000 ha⁻¹ yr⁻¹, considering the increased crop yields and reduced fertilizer costs. Similarly, a study by Galinato *et al.* (2011) reported that the application of wood biochar at 10 t ha⁻¹ in a U.S. wheat-pea rotation had a benefit-cost ratio of 1.7, indicating a profitable investment.

However, other studies have found that the economic returns of biochar application may be marginal or negative, depending on the specific conditions (Clare *et al.*, 2014; Dickinson *et al.*, 2015). For example, a study by Clare *et al.* (2014) estimated that the application of sugarcane bagasse biochar at 10 t ha⁻¹ in an Australian sugarcane farm would result in a net economic loss of \$60 ha⁻¹ yr⁻¹, considering the high cost of biochar and the limited yield benefits. Similarly, a study by Dickinson *et al.* (2015) found that the economic feasibility of biochar application in a Canadian canola-wheat rotation was highly sensitive to the biochar price and the yield response, with break-even prices ranging from \$90 to \$500 t⁻¹, depending on the scenario.

8.2. Feedstock availability and logistics

Another critical challenge for the large-scale adoption of biochar in integrated farming is the availability and logistics of feedstock supply (Mašek *et al.*, 2019). The type, quantity, and quality of feedstock can significantly influence the properties and performance of biochar as a soil amendment (Singh *et al.*, 2010). Therefore, ensuring a reliable and sustainable supply of appropriate feedstock is essential for the success of biochar projects (Sohi *et al.*, 2010).

The availability of biochar feedstock varies widely across regions and depends on factors such as land use, biomass productivity, and competing uses (Woolf *et al.*, 2010). In general, agricultural and forestry residues, such as crop straws, wood chips, and sawdust, are the most abundant and low-cost feedstocks for biochar production (Sohi *et al.*, 2010). However, the use of these residues for biochar may compete with other applications, such as animal feed, bioenergy, or soil conservation, which can limit their availability (Woolf *et al.*, 2010). In some cases, dedicated biomass crops, such as miscanthus or switchgrass, may be grown specifically for biochar production, but this can raise concerns about land use change and food security (Mašek *et al.*, 2019).

The logistics of feedstock supply, including collection, storage, and transportation, can also pose significant challenges for biochar projects (Mašek *et al.*, 2019). The bulky and low-density nature of most biomass feedstocks can make their transportation over long distances economically and environmentally unsustainable (Stelte *et al.*, 2012). Therefore, biochar production facilities need to be located close to the feedstock sources to minimize transportation costs and emissions (Meyer *et al.*, 2011). However, this can limit the scale and location of biochar projects, particularly in regions with dispersed or seasonal feedstock availability (Shackley *et al.*, 2015).

8.3. Standardization and quality control of biochar products

One of the main challenges for the widespread adoption of biochar in integrated farming is the lack of standardization and quality control of biochar products (Lehmann & Joseph, 2015). The properties and performance of biochar can vary widely depending on the feedstock type, production conditions, and post-treatment processes, which can lead to inconsistent results and uncertainties for end-users (Sohi *et al.*, 2010). Therefore, establishing clear and harmonized standards and quality control procedures for biochar production and application is crucial for ensuring its safety, effectiveness, and reliability as a soil amendment (Tammeorg *et al.*, 2017).

Several initiatives have been undertaken to develop biochar standards and certification systems at the national and international levels (Meyer *et al.*, 2017). For example, the International Biochar

Initiative (IBI) has developed a voluntary biochar certification program based on the IBI Biochar Standards, which specify the minimum criteria for biochar quality, including chemical properties, toxicant concentrations, and sustainability requirements (IBI, 2015). Similarly, the European Biochar Certificate (EBC) provides a voluntary certification system for biochar products based on a set of quality and sustainability criteria, including feedstock eligibility, production process, and product properties (EBC, 2012). Other countries, such as the USA, Australia, and Japan, have also developed their own biochar standards and certification schemes (Meyer *et al.*, 2017).

However, the adoption and harmonization of biochar standards and certification systems face several challenges, such as:

- **Variability of biochar properties:** The wide range of potential feedstocks and production conditions can result in biochars with diverse physical, chemical, and biological properties, which can be difficult to standardize and compare (Singh *et al.*, 2010). Therefore, biochar standards need to be flexible enough to account for this variability, while still ensuring a minimum level of quality and safety (Lehmann & Joseph, 2015).
- **Lack of consensus on key biochar properties:** There is still a lack of scientific consensus on which biochar properties are most important for its performance as a soil amendment, and how these properties should be measured and reported (Tammeorg *et al.*, 2017). This can lead to inconsistencies and confusion among biochar producers, users, and regulators (Meyer *et al.*, 2017).
- **Limited market demand and incentives:** The adoption of biochar standards and certification systems can be hindered by the limited market demand and incentives for certified biochar products (Jirka & Tomlinson, 2013). Many farmers and end-users may not be aware of or willing to pay a premium for certified biochar, which can discourage producers from investing in the certification process (Guo *et al.*, 2016).
- **Regulatory and policy barriers:** The lack of clear and supportive policies and regulations for biochar production and use can also pose barriers to the adoption of biochar standards and certification systems (Vochozka *et al.*, 2016). In many countries, biochar is still not recognized as a distinct product category, which can create uncertainties and obstacles for its standardization and quality control (Meyer *et al.*, 2017).

8.4. Policy support and incentives for biochar use

The adoption and upscaling of biochar use in integrated farming systems can be greatly influenced by the policy support and incentives provided by governments and other stakeholders (Liu & Zhang, 2021). Policy instruments, such as regulations, subsidies, taxes, and market-based

mechanisms, can play a crucial role in creating an enabling environment for biochar production and use, by addressing the economic, social, and environmental barriers and opportunities associated with this technology (Joseph *et al.*, 2013).

8.5. Knowledge gaps and future research directions

Despite the growing body of research on biochar and its potential applications in integrated farming systems, several knowledge gaps and uncertainties remain, which need to be addressed through further research and development (Tammeorg *et al.*, 2017). Some of the key knowledge gaps and future research directions for biochar in integrated farming systems include:

- Long-term effects of biochar on soil properties and crop productivity: While many studies have reported positive short-term effects of biochar on soil health and crop yields, the long-term impacts of biochar application are still poorly understood (Sohi *et al.*, 2010). More long-term field experiments and monitoring studies are needed to assess the persistence, stability, and resilience of biochar-induced changes in soil properties and crop performance, across different soil types, cropping systems, and environmental conditions (Tammeorg *et al.*, 2017).
- Interactions between biochar and other soil management practices: Biochar is often applied in combination with other soil management practices, such as fertilization, irrigation, tillage, or crop rotation, which can influence its effects on soil and crop outcomes (Joseph *et al.*, 2013). More research is needed to understand the synergies, trade-offs, and interactions between biochar and other soil management practices, and to optimize the integration of biochar into existing farming systems and practices (Agegnehu *et al.*, 2017).
- Mechanisms of biochar-induced changes in soil biota and ecosystem functions: Biochar has been shown to influence the abundance, diversity, and activity of soil microorganisms, fauna, and flora, which play critical roles in soil health, nutrient cycling, and ecosystem services (Lehmann *et al.*, 2011). However, the underlying mechanisms and implications of biochar-induced changes in soil biota and ecosystem functions are still not fully understood (Thies *et al.*, 2015). More research is needed to elucidate the complex interactions between biochar, soil biota, and ecosystem processes, and to assess the potential benefits and risks of biochar for soil biodiversity and ecosystem resilience (Tammeorg *et al.*, 2017).

CHAPTER - 9

Case studies

To further illustrate the potential and diversity of biochar applications in integrated farming systems, this section presents several case studies of successful biochar projects from around the world, showcasing different feedstocks, production methods, application strategies, and outcomes.

9.1. Small-scale farms in developing countries

Biochar has been shown to have significant potential for improving soil fertility and crop productivity in small-scale farms in developing countries, where soil degradation and food insecurity are major challenges (Gwenzi *et al.*, 2015). One successful example is the BioFarm Project in Cambodia, which has been working with smallholder farmers to produce and apply biochar from rice husks and other agricultural residues (BioFarm, 2021).

9.2. Large-scale commercial farms in developed countries

Biochar has also been successfully applied in large-scale commercial farms in developed countries, where it can provide multiple benefits for soil health, crop productivity, and environmental sustainability (Joseph *et al.*, 2013). One notable example is the Pacific Pyrolysis project in Australia, which has been producing and applying biochar from woody biomass in a large-scale commercial orchard since 2014 (Pacific Pyrolysis, 2021).

9.3. Agroforestry and silvo-pastoral systems

Biochar has also shown promise for enhancing soil health and productivity in agroforestry and silvo-pastoral systems, where trees and crops or livestock are integrated on the same land for multiple benefits (Stavi *et al.*, 2016). One successful example is the Biochar for Sustainable Soils (B4SS) project in Kenya, which has been promoting the use of biochar in smallholder agroforestry systems since 2014 (B4SS, 2021).

9.4. Organic and conservation agriculture

Biochar has also been explored as a promising soil amendment for organic and conservation agriculture systems, which aim to promote soil health, biodiversity, and sustainability by minimizing soil disturbance, maximizing soil cover, and diversifying crop rotations (Giller *et al.*, 2015). One notable example is the European Biochar for Sustainable Agriculture (EBSA) project, which has been investigating the potential of biochar for enhancing

soil quality and crop performance in organic and conservation agriculture across several European countries (EBSA, 2021).

CHAPTER 10

Conclusions and recommendations

This review article has explored the multifaceted roles of biochar in enhancing soil health and crop productivity in integrated farming systems. The key findings of this review can be summarized as follows:

- Biochar is a carbon-rich material produced by the pyrolysis of biomass, which has unique physical, chemical, and biological properties that can improve soil quality and fertility, such as high surface area, porous structure, alkaline pH, high cation exchange capacity, and stable carbon content.
- Biochar can be produced from a wide range of feedstocks, such as crop residues, wood waste, and animal manure, using different pyrolysis technologies, such as slow pyrolysis, fast pyrolysis, and gasification, which influence the yield, quality, and properties of the resulting biochar.
- Biochar application can enhance soil physical properties, such as soil structure, water holding capacity, and bulk density, which can improve soil aeration, drainage, and root growth, as well as reduce soil erosion and compaction.
- Biochar application can enhance soil chemical properties, such as soil pH, cation exchange capacity, and nutrient retention, which can increase soil fertility, nutrient availability, and nutrient use efficiency, as well as reduce nutrient leaching and greenhouse gas emissions.
- Biochar application can enhance soil biological properties, such as microbial biomass, diversity, and activity, which can improve soil health, nutrient cycling, and disease suppression, as well as stimulate beneficial soil-plant interactions, such as mycorrhizal symbiosis and plant growth-promoting rhizobacteria.
- Biochar application can improve crop growth and productivity, by enhancing nutrient uptake, water use efficiency, and stress tolerance, as well as reducing the incidence of pests and diseases, depending on the biochar type, application rate, and crop system.
- Biochar application can contribute to climate change mitigation and adaptation, by sequestering carbon in soils, reducing greenhouse gas emissions, and improving soil and crop resilience to drought, heat, and other climate stresses, as well as providing opportunities for carbon farming and ecosystem services.

- Biochar adoption in integrated farming systems faces several challenges and opportunities, such as economic feasibility, feedstock availability, quality control, policy support, and knowledge gaps, which require a holistic and collaborative approach to research, development, and implementation of biochar systems.
- Successful case studies of biochar application in integrated farming systems, such as small-scale farms in developing countries, large-scale commercial farms in developed countries, agroforestry and silvo-pastoral systems, and organic and conservation agriculture, demonstrate the potential and diversity of biochar benefits and synergies in different contexts and scales.

CHAPTER 11

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